

AIMS Environmental Science, 12(2): 352-372.

DOI: 10.3934/environsci.2025016

Received: 19 October 2024 Revised: 31 March 2025 Accepted: 03 April 2025 Published: 15 April 2025

https://www.aimspress.com/journal/environmental

Research article

Assessment of indoor air quality in Tunisian childcare establishments

Meher Cheberli^{1,2,3}, Marwa Jabberi^{1,2,4}, Sami Ayari⁵, Jamel Ben Nasr⁶, Habib Chouchane², Ameur Cherif², Hadda-Imene Ouzari^{5,*} and Haitham Sghaier^{1,2,*}

- ¹ Laboratory Energy and Matter for Development of Nuclear Sciences LR16CNSTN02, National Center for Nuclear Sciences and Technology, Sidi Thabet Technopark, 2020, Tunisia
- ² Univ. Manouba, ISBST, BVBGR-LR11ES31, Biotechpole Sidi Thabet, 2020, Ariana, Tunisia
- ³ Ministry of Health, Bab Saadoun 1006 Tunis, Tunisia
- ⁴ Biochemistry and Molecular Biology Lab of Faculty of Sciences, Risks Related to Environmental Stress, Struggle and Prevention (UR17ES20), Bizerte, Zarzouna, University of Carthage, Tunisia
- ⁵ Laboratory of Microorganisms and Active Biomolecules LR03ES03, Department of Biology, Faculty of Science, University of Tunis El Manar, 2092 Tunis, Tunisia
- National Agronomic Institute of Tunis, Tunisia 43 Av. Charles Nicolle, Tunis 1082, University of Carthage, Tunisia
- * Correspondence: Hadda-Imene Ouzari: imene.ouzari@fst.utm.tn, Tel: 21698925872; Haitham Sghaier: haitham.sghaier@cnstn.rnrt.tn, haitham.sghaier@supcom.tn, Tel: 21652058289.

Abstract: Maintaining healthy indoor air quality (IAQ) in childcare settings is essential for infants and young children, as it directly impacts their early learning, development, and overall well-being. Given their vulnerability, continuous IAQ monitoring in these environments is crucial to ensuring a safe and supportive atmosphere. This study aimed to assess IAQ factors that may affect occupant health by measuring indoor concentrations of particulate matter (PM₁₀), selected gases such as carbon dioxide (CO₂) and formaldehyde (CH₂O), and thermal conditions including temperature and relative humidity. Additionally, airborne microorganism levels were analyzed, and potential environmental factors influencing microbial abundance were investigated in three childcare centers in Megrine, Tunisia, across three seasonal periods. Results revealed frequent occurrences of hygrothermal discomfort and elevated levels of CO₂, CH₂O, and PM₁₀, particularly in overcrowded classrooms with poor ventilation and heating. Pathogenic bacterial species, including *Staphylococcus epidermidis*, *Staphylococcus haemolyticus*, *Bacillus cereus*, and *Bacillus licheniformis*, were repeatedly detected. Significant correlations were found between bacterial abundance and

environmental factors such as PM₁₀, CO₂ levels, temperature, and humidity. These findings provide valuable insights into IAQ dynamics in childcare environments, highlighting the need for improved ventilation and air quality management strategies to safeguard children's health and well-being.

Keywords: airborne bacteria; childcare facilities; indoor air quality; metataxonomic analysis; particulate matter; physicochemical parameters

1. Introduction

The air inside enclosed spaces is a mixture of physical, chemical, and biological pollutants, which originate from outside air, materials, combustion devices, and human activities. The indoor air quality (IAQ) of homes, offices, schools, or other buildings is an essential determinant of healthy living and well-being of people [1]. Most people spend 80%–95% of their time in indoor environments with an average of 10–14 m³ of air per day [2]. In most studies, symptoms such as dizziness, headache, nausea, and irritation of the eyes, nose, and throat have been shown to be linked to poor IAQ [3].

Indoor air quality has been the focus of numerous studies due to growing concerns within the scientific community regarding its impact on occupant health and comfort [4–11]. However, poor IAQ in schools can be particularly severe compared to other types of buildings, primarily due to higher occupant density. Elevated contaminant levels may stem from various factors, including the intrusion of outdoor pollutants, the building's physical state, cleaning practices, and the effectiveness of the ventilation system [12–17].

The World Health Organization (WHO) has selected particulate matter (PM_{2.5} and PM₁₀) and some gaseous compounds as crucial for checking IAQ, namely radon, carbon monoxide, nitrogen dioxide, polycyclic aromatic hydrocarbons (PAH), formaldehyde (CH₂O), and other volatile organic compounds [18]. Formaldehyde, classified as a certain carcinogen for humans, represents a priority pollutant in school buildings due to excessive and regular use of school supplies (glue, gouache, ink, etc.) and cleaning products (detergents, air fresheners, etc.) representing potential sources of CH₂O emissions in classrooms. Several studies have demonstrated that indoor exposure to formaldehyde has been associated with respiratory and asthma symptoms and decreased lung function in children [19]. On the other hand, hygrothermal comfort and air humidity in classrooms are also relevant in the study of IAQ since they affect the perceived comfort of IAQ, thus causing symptoms of eye and respiratory irritation and impacting children's performance [20]. Although not considered a pollutant per se in indoor environments, carbon dioxide (CO₂) has been used as a relevant indicator of adequate ventilation in classrooms. Studies on children have shown that increased CO₂ concentration in the classroom decreases school attendance over a short period of time [21–40], while persistent exposure to PM₁₀ could disrupt children's lung development later in life [41].

In many environments, particularly in daycare settings, bioaerosols can disrupt normal activities. Exposure to these airborne particles, which contain microorganisms and their byproducts, may lead to respiratory issues and other health complications such as toxic reactions and infections [42,43]. Airborne microorganisms, such as bacteria, fungi, and yeast, generally come from humans but also from animals, plants, soil, building materials, and the external environment [44]. In

schools, children tend to have high activity levels, which generally leads to higher levels of airborne microorganisms. Epidemiological studies show that too high a concentration of microorganisms in the air can be allergenic and that sometimes even very low concentrations of particular microorganisms can cause serious illnesses [45]. Additionally, the amount of microbes present in indoor school air has a direct impact on students' mental health, physical development, and performance [46]. Diriba et al. revealed that failure to clean and check heating and air conditioning systems can allow microbial growth, causing rhinitis, bronchitis, pharyngitis, pneumonia, conjunctivitis, and keratitis [47]. Bacteria in indoor environments primarily originate from human sources, including skin, the oral cavity, intestines, and clothing [48]. Some airborne bacteria are toxic, allergenic, or infectious, posing health risks such as respiratory and dermatological infections [49,50].

Given the health risks associated with poor IAQ, its assessment is crucial for effective risk management. In many countries, IAQ monitoring in childcare facilities and schools has become a legal requirement. This study aims to investigate the physicochemical and microbiological quality of indoor air in three childcare centers in Tunisia and to explore correlations between environmental parameters and bacterial load.

2. Materials and methods

2.1. Site introduction

This study was conducted in three collective childcare establishments located in the town of Megrine, northern Tunisia: a kindergarten (E1) and two daycare schools (E2 and E3). To perform the study, one room from each establishment was selected: RE1 in E1, RE2 in E2, and RE3 in E3. Data collection took place over three distinct periods: summer (early October 2020), winter (late December 2020), and spring (April–May 2021) (Table 1).

Site	Room	Class/grade	Floor	Occupation by children		
		(years)		Number of children	Total occupancy period (minutes)	
RE1	Activity room	3–4	Ground floor	12 (summer)	273 (summer)	
				20 (winter)	222 (winter)	
				17 (spring)	184 (spring)	
RE2	Duty room	6–7	Ground floor	20 (summer)	271 (summer)	
				20 (winter)	260 (winter)	
				19 (spring)	170 (spring)	
RE3	Duty room	5–6	First floor	14 (summer)	240 (summer)	
				16 (winter)	285 (winter)	
				15 (spring)	255 (spring)	

Table 1. Main characteristics of each studied site.

2.2. IAQ characterization

During the children's normal activities (full rooms), comfort parameters [temperature (T) and

relative humidity (RH)] and CO_2 concentrations were measured inside the rooms using an analyzer IAQ (model Q-TRAK 7575). Formaldehyde (CH₂O) concentrations were measured using a formaldehyde meter (model HAL-HFX205). Aerosol monitors (DUSTTRAK II model) were employed to simultaneously measure particulate matter with an aerodynamic diameter of less than 10 μ m (PM₁₀) inside the rooms and outside (PM10_{ext}) of the selected spaces. All measurements were conducted at regular intervals of 1 min.

Comfort conditions related to air temperature and relative humidity in the rooms were assessed using a hygrothermal comfort diagram [51]. This diagram defines a comfort zone based on ambient temperature and corresponding humidity levels perceived as comfortable. Two discomfort zones are identified: a *humid zone* with elevated temperature and humidity, and a *dry zone* characterized by lower temperature and humidity levels.

2.3. Metataxonomic analysis

In conjunction with these measurements, air samples designated for microbiological analyses were collected within rooms RE2 and RE3 during the summer and winter periods using a bio-collector (model BK-BAS). Next-generation sequencing (NGS) technology was employed to explore the diversity of airborne microorganisms and pathogenic bacteria within these environments.

DNA extraction was conducted utilizing the GeneJET Genomic DNA Purification kit (Thermo Scientific) following the manufacturer's instructions. Metagenomic sequencing library preparation was carried out according to the Illumina protocol (Illumina Inc., San Diego, CA, USA), targeting the V3-V4 region of the 16S rDNA gene, employing universal primer sets as follows: forward = 5' (CCTACGGGNGGCWGCAG) and reverse = 5' (GACTACHVGGGTATCTAATCC) [52]. DNA concentration was quantified using a Qubit fluorometer (Invitrogen, USA). Amplification was performed in a 25 µL mixture, incorporating KAPA HiFi Hot-Start PCR kit (Kapa Biosystems), 10 mM of each primer, and 5 ng of DNA. PCR conditions comprised an initial step at 95 °C for 3 min followed by 25 cycles (95 °C for 30 s, 52 °C for 30 s, 55 °C for 30 s, 72 °C for 30 s), with a final extension at 72 °C for 5 min. Following PCR, amplicons underwent purification using magnetic beads with two 80% ethanol washes. Subsequently, a second PCR was conducted, integrating specific adapters and indexes for sample identification and employing the KAPA HiFi Hot-Start PCR kit. Nuclease-free water was utilized to adjust the final volume to 50 µL. Following dilution and normalization of the combined library, adhering to Illumina's recommendations for sequencing on Miseq, the sequencing process involved denaturation with 0.2 N NaOH, further dilution to 4 pM in sequencing buffer (HT1), and loading into the Miseq cartridge for sequencing. The raw data underwent analysis using the One Codex data platform due to its various advantages [53,54].

2.4. Data treatment

For the CO₂ data acquired, comparisons were made with the American Society of Heating, Refrigerating and Air-Conditioning Engineers (ASHRAE) standard reference value of 1000 ppm [55]. Formaldehyde and PM₁₀ data were evaluated against the World Health Organization (WHO) recommended values of 100 μg/m³ for CH₂O and 50 μg/m³ for PM₁₀ [56]. Moreover, correlations between physicochemical parameters and the bacterial community were established

utilizing correlograms generated via the FaDA application [57].

3. Results

3.1. Occupancy and ventilation conditions

During the summer period, the highest occupancy rates were observed in the study rooms. In winter and spring, room RE1 had the highest density, with 1.57 m² and 1.8 m² per child, respectively (Table 2). Room RE3 had the smallest surface area of external openings (windows, doors, and patio doors), with only 0.81 m² available (compared to the required 5.25 m²), while room RE2 had the largest opening area, measuring 4.55 m² (compared to the required 7.93 m²).

Table 2. Occupancy rates and surfaces of the openings in studied rooms.

D	Area (m ²)	Occupancy	rate (m ² per	child)	Surface of the openings		
Room		Summer	Winter	Spring	Average	$s_1(m^2)^a$	$s_2 (m^2)^b$
RE1	31.5	2.62	1.57	1.8	1.99	2.93	5.25
RE2	47.6	2.38	2.38	2.5	2.42	4.55	7.93
RE3	31.4	2.24	1.96	2.09	2.09	0.81	5.23

Notes: as₁ represents the surface of the openings (m²) available in the room; bs₂ represents the minimum surface area of openings (m²) imposed by French thermal regulations: one-sixth of the surface area of the room.

3.2. Characterization of IAQ

3.2.1. Thermal comfort parameters

The hygrothermal comfort conditions within the classrooms varied significantly by season. During the summer, all three rooms were situated outside the hygrothermal comfort zone, characterized by hot and humid environments. In winter, rooms RE2 and RE3 entered the comfort zone, while RE1 remained outside. In spring, all rooms exhibited favorable hygrothermal conditions (Figure 2). Studies have shown that inadequate hygrothermal conditions increase the risk of indoor contamination by bacteria, viruses, fungi, and mites, which can lead to health issues such as rhinitis, allergies, asthma, and in severe cases, respiratory and lung infections [58–60].

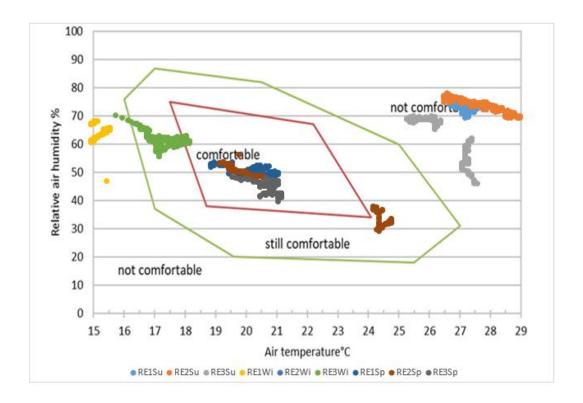


Figure 1. Hygrothermal comfort zones of classrooms (RE1, RE2, and RE3) during the summer (Su), winter (Wi), and spring (Sp) periods.

3.2.2. CO₂ concentrations

The highest average CO₂ concentrations were recorded in the winter, with room RE1 showing a mean value of 3186.6 ppm (range: 1891–4394 ppm) and room RE3 a mean value of 2781.6 ppm (range: 2059–3433 ppm) (Figure 2). All CO₂ measurements in RE3 (in both winter and spring) and RE1 (in winter) exceeded the ASHRAE-recommended threshold of 1000 ppm (Figure 3). These findings are consistent with current literature, which highlights high CO₂ levels as a key concern in classroom air quality [61–68].

Correlation analysis of the monitored parameters revealed a strong positive correlation between CO₂ levels and indoor PM₁₀ concentrations, with correlation coefficients of 0.90, 0.80, and 0.58 in RE1, RE3, and RE2, respectively. These results align with the study by Simoni and Annesi-Maesano [69], which investigated IAQ in 46 classrooms across 21 schools in six cities in five European countries and found significant correlations between CO₂ and PM₁₀ levels (r=0.66) [70]. Nevertheless, strong negative correlations were observed between CO₂ concentrations and air temperature levels, with coefficients of the order of -0.90, -0.84, and -0.80 in RE3, RE1, and RE2, respectively. Similarly, findings from a study conducted in 310 schools and nurseries across France between 2009 and 2011 also noted significant negative correlations between CO₂ levels and temperature [70].

The elevated CO₂ concentrations observed in RE1 and RE3 during winter and spring were primarily attributed to inadequate ventilation and higher room occupancy compared to other spaces. These rooms, RE1 and RE3, had the smallest natural ventilation openings, and throughout these

seasons, they experienced the highest levels of crowding.

Furthermore, windows remained closed throughout the school day in RE1 during winter and RE3 during both winter and spring. These findings suggest a clear association between inadequate ventilation and increased CO₂ levels in classrooms.

In contrast, room RE2, which had the best ventilation and occupancy conditions, recorded the lowest CO₂ concentrations (Figures 2 and 3, Table 2). A study by Clausen et al. (2014) in 785 Danish classrooms supported these findings, demonstrating that ventilation during breaks, such as opening windows, reduced the proportion of classrooms with CO₂ concentrations above 1000 ppm from 60% to 39% annually, compared to rooms where windows were never opened [71].

Conversely, during winter, combustion heating was used in room RE3 throughout the school day without adequate ventilation, likely contributing to the excessive increase in CO₂ concentrations. These results align with those of Hanoune and Carteret (2015), who demonstrated that the operation of individual combustion devices can elevate CO₂ concentrations up to 4500 ppm. Their study on the indoor air quality of seven homes equipped with combustion heaters revealed that CO₂ concentrations above 1000 ppm were consistently attributed to combustion sources [72]. Similarly, a study conducted in Croatia during the heating season found that CO₂ concentrations in 60 classrooms, all of which were poorly ventilated, surpassed international guidelines. The average CO₂ concentrations in these rooms ranged between 862 and 2415 ppm [73]. These results indicate that CO₂ concentrations above 1000 ppm are relevant indicators of deficient ventilation flow in classrooms.

3.2.3. CH₂O concentrations

The highest average concentrations of formaldehyde (CH₂O) were detected during summer in rooms RE3 and RE2, with mean values of 1115.8 μg/m³ (range: 0–6385.6 μg/m³) and 650.8 μg/m³ (range: 0–13606.2 μg/m³), respectively (Figure 3). In these rooms, the rate of exceedance of the WHO-recommended limit value of 100 μg/m³ was particularly high in summer, reaching 65% in RE3 and 33% in RE2 (Figures 4 and 5). These results are consistent with recent studies that have also reported elevated levels of CH₂O in classrooms [66,74–77]. The presence of high CH₂O levels is primarily attributed to the widespread use of hydroalcoholic gels and solutions for hand disinfection, as indicated by the work of Santos Catai et al. [78], as well as surface and floor cleaning procedures implemented in the classrooms as part of the actions undertaken by these establishments to fight against the SARS-CoV-2 virus during the COVID-19 pandemic.

However, it is important to note that the electrochemical sensor used for CH₂O measurement (HAL-HFX05) has known cross-sensitivities, particularly to ethanol (10%) and isopropanol (2%), both of which are key components of hand disinfectants. While our observations align with previous studies indicating ethanol oxidation as a potential secondary source of CH₂O formation in indoor environments [78], further investigation using alternative analytical techniques, such as colorimetric assays or gas chromatography, would help refine these measurements and minimize potential interference.

3.2.4. PM₁₀ particles

As illustrated in Figure 6, PM_{10} concentrations reached their highest levels in rooms RE3 and RE1 during the winter, with values of 114 μ g/m³ (range: 85–138 μ g/m³) and 85 μ g/m³ (range: 69–116 μ g/m³), respectively. During this season, PM_{10} measurements in these rooms exceeded the WHO-recommended limit of 50 μ g/m³ (Figure 7). During the summer, the highest median PM_{10} concentrations were observed in RE2 (mean: 68 μ g/m³, range: 48–180 μ g/m³) and RE3 (mean: 67 μ g/m³, range: 46–91 μ g/m³), with 95% and 91% of readings in these rooms surpassing the WHO limit.

Correlation analyses between indoor and outdoor PM_{10} concentrations revealed significant associations, with correlation coefficients of 0.45, 0.80, and 0.83 in RE1, RE3, and RE2, respectively. These results are similar to those found by Matic et al. [79] who concluded that the increase in PM_{10} concentrations inside eight naturally ventilated schools in Serbia (among nine studied schools) was significantly affected by the increase in PM_{10} emissions from outside, with correlation coefficients varying between 0.45 and 0.95.

However, an analysis of the indoor-to-outdoor (I/O) ratios of PM₁₀ concentrations indicated that the dominant sources of PM₁₀ varied by season. In RE1 and RE2, the median I/O ratios were greater than 1 during both summer and winter (1.42 and 2.41 in RE1; 1.32 and 1.31 in RE2), suggesting that indoor sources were predominant. According to several studies, children's activities, movement, and the presence of furniture such as whiteboards, tables, and chairs are key indoor sources of particulate emissions in classrooms [80,81]. A study by Oliveira et al. [82] showed that air conditioning, heating systems, and cleaning activities also contribute significantly to indoor PM₁₀ levels in school environments.

In contrast, during the spring, the I/O ratios in these same rooms approached 1, indicating a shift toward the predominance of outdoor sources of PM₁₀. In RE3, the contribution of indoor sources increased progressively over the seasons. The I/O ratio in this room was approximately 1 during the summer, increased to 1.38 in winter, and peaked at 4.9 in spring. The significant rise in PM₁₀ levels inside RE3 during winter can largely be attributed to the operation of the heating system, as supported by numerous studies that have shown the role of heating in elevating particle emissions in enclosed spaces [83,84].

A model-based study by Hong et al. [85] showed a strong positive association between the size of window openings in buildings and indoor PM_{10} concentrations. These authors concluded that increasing window openings improves indoor ventilation and thus reduces PM_{10} concentrations inside buildings. This could partly explain the presence of PM_{10} at relatively high levels in RE3 (all seasons) and RE1 (in winter). These rooms, although less exposed to the effect of outside air pollution, nevertheless had the smallest opening surfaces, with s_1/s_2 ratios of around 0.31 and 1.12, respectively (Figures 6 and 7 and Table 2).

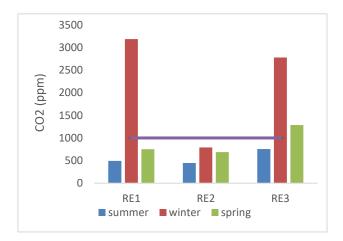


Figure 2. Average concentrations of CO₂ by room. The line at 1000 ppm indicates the ASHRAE standard reference value.

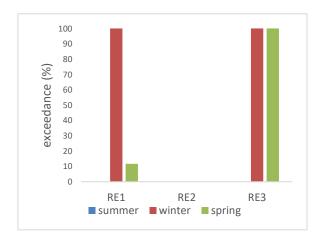


Figure 3. Rate of CO₂ concentrations exceeding the ASHRAE reference value by room.

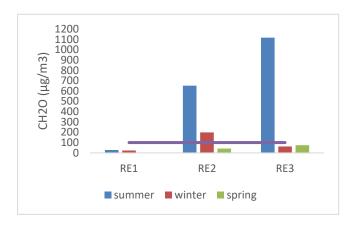


Figure 4. Average concentrations of CH_2O by room. The line at $100 \ \mu g/m^3$ indicates the value recommended by the WHO.

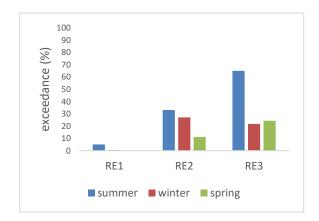


Figure 5. Rate of CH₂O concentrations exceeding the limit value recommended by the WHO according to the room.

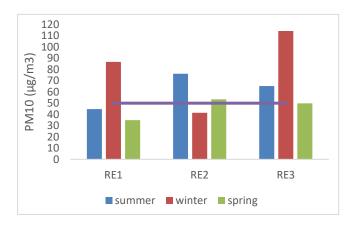


Figure 6. Average concentrations of PM_{10} by room. The line at 50 $\mu g/m^3$ indicates the value recommended by the WHO.

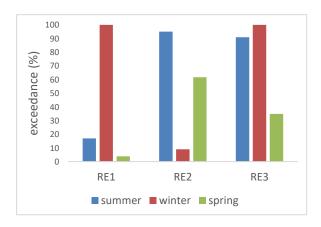


Figure 7. Rate of PM_{10} concentrations exceeding the limit value recommended by the WHO according to the room.

3.3. Metataxonomic analysis

3.3.1. Bacterial species in classrooms

The accession number for the Short Read Archive (SRA) data generated in this study is PRJNA1059615. Bacillus and Staphylococcus, each comprising four species, represented the bacterial genera most detected, both in terms of frequency and abundance (Table 3). A total of 24 species of airborne bacteria were identified from the collected air samples, with detection frequencies ranging from one to three occurrences. Sixteen distinct bacterial species were detected only once, while eight species were observed at least twice. In terms of pathogenicity, certain species classified as pathogenic to humans [86] were detected in the air samples collected, namely Staphylococcus epidermidis (three occurrences), Bacillus cereus (three occurrences), and Staphylococcus haemolyticus (two occurrences), as well as Bacillus licheniformis, Bordetella petrii, Proteus mirabilis, Klebsiella aerogenes, Vagococcus fluvialis, Providencia rittgeri, and Pseudomonas fluorescens (each a single occurrence). Among these pathogens, six species were detected simultaneously in the air sample collected from RE2 in summer and four species in the sample collected in summer from RE3 (Table 3). Regarding abundance, B. licheniformis and B. petrii were detected with the highest relative abundance rates in the air sample collected from RE2 in summer, reaching 29.61% and 8.8%, respectively. As for S. epidermidis and B. cereus, their abundances reached a maximum in air samples collected in winter from RE3 (5.86%) and RE2 (8.14%), respectively.

It should also be noted that other bacterial species classified as risk class 2 pathogens (microorganisms that can cause disease in humans and constitute a danger for people directly exposed to them) have been found more than once in the air samples collected [86]. These mainly include *Staphylococcus saprophyticus* and *Bacillus cereus* (each three occurrences) and *Staphylococcus hominis* (twice) (Table 3).

Table 3. Average values of physicochemical parameters, relative abundances of isolated bacterial species, and significant correlations found between them in the indoor air of rooms RE2 and RE3 in the summer (Su) and winter (Wi) periods.

				` ' -		
Room		RE2Su	RE2Wi	RE3Su	RE3Wi	Significant correlations ^C
Physicochemical parameters						
Temperature (T, °C)		27.75	11.69	26.35	17.4	
Relative humidity (RH, %)		73.71	54.8	64.28	61.2	
Carbone dioxide (CO ₂ , ppm)		445.97	791.8	753.67	2781.53	
Formaldehyde (CH ₂ O, μg/m ³)	650.8	198.7	1115.79	62.74		
Indoor concentration of PM ₁₀ (PM	$(10, \mu g/m^3)$	76.19	41.47	65.22	114.34	
Outdoor concentration of PM ₁₀ (Pl	M_{10ext} , $\mu g/m^3$)	65.88	33.05	69.77	102.63	
Bacterial community						
Bacterial species	Frequency	% of clas	sified reads			
Bacillus cereus*	3	1.84	8.14	3.35	-	PM ₁₀ (r=-0.92, P=0.0734) ^b
						PM _{10ext} (r=-0.94, P=0.058) ^b
Bacillus firmus	3	2.42	3.74	-	3.43	CH_2O (r=-0.95, P =0.0405)
Staphylococcus epidermidis*	3	1.41	3.01	-	5.86	CH ₂ O (r=-0.92, P=0.0757) ^b
Staphylococcus saprophyticus	3	-	2.63	4.18	12.52	CO_2 (r=0.97, P =0.021) ^a
						R.oc (r=-0.97, P=0.032) ^a
Pseudomonas stutzeri	3	-	7.08	2.19	3.13	RH (r=-0.95, P=0.0404) ^a
						T (r=-0.93, P=0.0681) ^b
Micrococcus luteus	3	-	2.28	5.22	1.05	-
Staphylococcus haemolyticus*	2	-	4.67	1.03	-	-
Staphylococcus hominis	2	1.79	-	1.5	-	T (r=0.95, P=0.049) ^a
Bacillus licheniformis*	1	29.61	-	-	-	-
Bacillus subtilis	1	11.79	-	-	-	-
Pseudomonas fluorescens*	1	-	-	-	2.32	-
Micrococcuss sp.	1	-	-	3.51	-	-
Bordetella petrii*	1	8.8	-	-	-	-
Proteus mirabilis*	1	3.31	-	-	-	-
Providencia rettgeri*	1	1.11	-	-	-	-
Leclercia adecarboxylata	1	-	1.43	-	-	-
Planococcus citreus	1	-	1.27	-	-	-
Sphingobacterium faecium	1	-	1	-	-	-
Klebsiella aerogenes*	1	-	-	5.64	-	-
Vagococcus fluvialis*	1	-	-	1.12	-	-
Psychrobacter faecalis	1	-	-	-	2.01	-
Psychrobacter celer	1	-	-	-	1.33	-
Acinetobacter johnsonii	1	-	-	-	1.01	-
Oceanobacillus profundus	1	-	-	-	1	-
Total number of species detected 24		9	10	9	10	-
Cumulative abundance of bacteria	of the genus	45.66	11.88	3.35	3.43	-

Bacillus (%)					
Cumulative abundance of bacteria of the genus	3.2	10.31	6.71	18.38	CO ₂ (r=0.94, P=0.0559) ^b
Staphylococcus (%)					

Notes: *Bacterial species classified as pathogenic to humans [86]. ^a Significant at the 0.05 level; ^b Significant at the 0.1 level. ^C Significant correlations found between the abundance of bacteria and physicochemical parameters

3.3.2. Correlations between physicochemical parameters and bacterial community

The correlation analysis conducted revealed that the diversity of staphylococci in the air samples is strongly correlated with the CO₂ concentration and the rate of occupancy by children. Indeed, a significant correlation was observed between the cumulative abundance of species of the *Staphylococcus* genus and the CO₂ concentration in the room (r=0.94, P=0.0559). In addition, the abundance of *Staphylococcus saprophyticus* showed a significant positive correlation with the CO₂ concentration in the room (r=0.97, P=0.021) and a significant negative correlation with the occupancy rate of the room by children (r=-0.97, P=0.032) (Table 3). These results support previous research highlighting the adaptability of *Staphylococcus* bacteria to environments with poor ventilation and high occupancy rates [87]. A study conducted by Madsen et al. [88] in different indoor spaces also showed that the diversity of staphylococci was significantly associated with the surface area per occupant. On the other hand, a significant positive correlation was observed between the abundance of *Staphylococcus hominis* and air temperature (r=0.95, P=0.049). In contrast, *Pseudomonas stutzeri* seems to favor relatively low temperatures and humidity.

The abundance of these species indeed shows significant negative correlations with air temperature (r=-0.93, P=0.0681) as well as relative humidity (r=-0.95, P=0.0404). Additionally, the abundances of *Bacillus firmus* (r=-0.95, P=0.0405) and *Staphylococcus epidermidis* are negatively correlated with the formaldehyde concentration (r=-0.95, P=0.0405).

4. Discussion of key findings and limitations

The findings of this study emphasize the importance of maintaining optimal IAQ in childcare environments, particularly given the potential adverse effects on children's health. Elevated concentrations of CO₂, CH₂O, and PM₁₀ were observed, especially in rooms with inadequate ventilation and higher occupancy. These findings are consistent with previous studies that have highlighted the role of poor ventilation in elevating CO₂ levels [61–68]. In our study, the highest CO₂ concentrations were recorded in RE1 and RE3 during winter and spring, which were both overcrowded and had limited ventilation openings. This suggests a direct link between overcrowding and ventilation insufficiencies, with clear implications for air quality and, by extension, occupant health.

The presence of formaldehyde (CH₂O) in the classrooms, particularly during summer, is another significant concern. The elevated CH₂O concentrations, which surpassed WHO's recommended limit in RE3 and RE2, were likely influenced by the widespread use of hydroalcoholic gels during the COVID-19 pandemic. Similar findings have been reported in other studies, indicating that cleaning and disinfecting protocols can inadvertently contribute to indoor air pollution [66,74–77]. The repeated exceedance of formaldehyde levels in these classrooms suggests the need for alternative

disinfection practices that minimize indoor air contamination.

Fine particulate matter (PM_{10}) levels also exceeded WHO's recommended limits, particularly in RE1 and RE3 during winter. Notably, the source of these particles appeared to be both indoor activities (e.g. heating and furniture) and outdoor pollution, with the latter becoming more significant in the spring. These findings align with previous research that suggests heating systems and activities like cleaning and movement contribute substantially to indoor PM_{10} concentrations [79,80].

A key finding of this study was the significant correlation between CO₂ concentrations and the abundance of certain bacterial species, including *Staphylococcus epidermidis* and *Bacillus licheniformis*. This suggests that elevated CO₂ levels may create an environment favorable to microbial growth, posing a potential health risk to children. The presence of these pathogenic bacteria in indoor air highlights the need for continuous monitoring and better control of both particulate matter and microorganisms to ensure the safety of childcare environments.

While these findings are valuable, this study is not without limitations. The small sample size (three childcare centers) and the seasonal variability in IAQ conditions may affect the generality of the results. Additionally, the study focused only on a limited number of pollutants and microbial species, and future research should explore other potential indoor contaminants. Moreover, variations in the type and effectiveness of air purification systems were not assessed, which could be an important area for future research.

This study emphasizes the need for improved ventilation and air quality management in childcare centers, particularly in regions with similar climate conditions. Simple interventions such as enhancing natural ventilation, reducing overcrowding, and employing effective air purification methods could significantly reduce the levels of CO₂, CH₂O, and PM₁₀, thereby mitigating health risks associated with poor IAQ.

5. Conclusions

The results of this study indicate that children in the examined classrooms experienced hygrothermal discomfort and were exposed to elevated levels of CO₂, PM₁₀, and formaldehyde. Additionally, microbiological culture-independent analyses revealed the presence of various bacterial species, including potential pathogens, highlighting a significant bioaerosol risk. The results emphasize that poor indoor environmental conditions such as inadequate ventilation, insufficient air exchange, overcrowding, and heating contribute to higher pollutant concentrations and microbial contamination. Consequently, implementing measures to optimize indoor air quality, regulate temperature and humidity, and improve classroom ventilation could significantly reduce children's exposure to airborne pollutants and bio-contaminants, ultimately fostering a healthier learning environment.

Use of AI tools declaration

The authors declare they have not used Artificial Intelligence (AI) tools in the creation of this article.

Acknowledgments

This research was conducted with the support of the Tunisian National Center for Nuclear Sciences and Technology (CNSTN), the Faculty of Sciences of Tunis (FST), and the Higher Institute of Biotechnology of Sidi Thabet (ISBST). We sincerely thank the Tunisian Ministry of Higher Education and Scientific Research, the common sequencing unit linked to the LR03ES03 at the FST and the Ministry of Health for their financial support, which was instrumental in carrying out this study.

Conflict of interest

The authors declare no conflicts of interest or personal relationships that could have influenced the work reported in this paper.

References

- 1. Faria T, Almeida-Silva M, Dias A, et al. (2016) Indoor air quality in urban office buildings. *Int J Environ Technol Manag* 19: 236–256. https://doi.org/10.1504/IJETM.2016.082243.
- 2. Awad AA, Farag SA (1999) An indoor bio-contaminants air quality. *Int J Environ Heal R* 9: 313–9. https://doi.org/10.1080/09603129973100.
- 3. Gomzi M, Bobic J (2009) Sick building syndrome: Do we live and work in unhealthy environment? *Period Biol* 111: 1. https://hrcak.srce.hr/35999.
- 4. Budd GM, Warhaft N (1966) Body temperature, shivering, blood pressure and heart rate during a standard cold stress in Australia and Antarctica. *J Physiol* 186: 216. https://doi.org/10.1113/jphysiol.1966.sp008030.
- 5. Heinzerling D, Schiavon S, Webster T, et al. (2013) Indoor environmental quality assessment models: a literature review and a proposed weighting and classification scheme. *Build Environ* 70: 210–222. https://doi.org/10.1016/j.buildenv.2013.08.027.
- 6. Lan L, Wargocki P, Wyon D.P, et al. (2011) Effects of thermal discomfort in an office on perceived air quality, SBS symptoms, physiological responses, and human performance. *Indoor Air* 21: 376–390. https://doi.org/10.1111/j.1600-0668.2011.00714.x.
- 7. Spengler JD, Sexton K (1983) Indoor air pollution: a public health perspective. *Science* 221: 9–17. https://doi.org/10.1126/science.6857273.
- 8. Sun YP, Zhu N (2012) Study on assessment of high temperature and humidity in working environment on human health. *Adv Mat Res* 610: 739–742. https://doi.org/10.4028/www.scientific.net/amr.610-613.739.
- 9. Vehvil€ainen T, Lindholm H, Rintam€aki H, et al. (2016) High indoor CO< sub> 2< /sub> concentrations in an office environment increases the transcutaneous CO< sub> 2< /sub> level and sleepiness during cognitive work. *J Occup Environ Hyg* 13: 19–29. https://doi.org/10.1080/15459624.2015.1076160.
- 10. Xue P, Mak CM, Ai ZT (2016) A structured approach to overall environmental satisfaction in high-rise residential buildings. *Energy Build* 116: 181–189. https://doi.org/10.1016/j.enbuild.2016.01.006_

- 11. Zhang X, Wargocki PZ (2011) Lian, Human responses to carbon dioxide, a follow-up study at recommended exposure limits in non-industrial environments. *Build Environ* 100: 162–171. https://doi.org/10.1016/j.buildenv.2016.02.014.
- 12. Pegas PN, Alves CA, Evtyugina MG, et al. (2011) Indoor air quality in elementary schools of Lisbon in spring. *Environ Geochem Health* 33: 455–468. https://doi.org/10.1007/s10653-010-9345-3.
- 13. Bradman A, Gaspar F, Castorina R, et al. (2012) Environmental exposures in early childhood education environments. Retrieved on 16 December 2022 from https://ww2.arb.ca.gov/sites/default/iles/classic/research/apr/past/-305.pdf
- 14. Branco PT, Nunes RA, Alvim-Ferraz MC, et al. (2015) Children's exposure to indoor air in urban nurseries—part II: gaseous pollutants' assessment. *Environ Res* 142: 662–670. https://doi.org/10.1016/j. envres. 2015. 08. 026
- 15. Branco P, Alvim-Ferraz MCM, Martins FG, et al. (2019) Quantifying indoor air quality determinants in urban and rural nursery and primary schools. *Environ Res* 176: 108534. https://doi.org/10.1016/j.envres. 2019. 108534
- 16. Hoang T, Castorina R, Gaspar F, et al. (2017) VOC exposures in California early childhood education environments. *Indoor Air* 27: 609–621. https://doi.org/10.1111/ina.12340
- 17. Zhang S, Mumovic D, Stamp S, et al. (2021) What do we know about indoor air quality of nurseries? A review of the literature. *Build Serv Eng Res Technol* 42: 603–632. https://doi.org/10.1177/01436244211009829
- 18. World Health Organization. Regional Office for E. (2010) WHO guidelines for indoor air quality: selected pollutants. Copenhagen: *World Health Organization. Regional Office for Europe* https://iris.who.int/handle/10665/260127.
- 19. Krzyzanowski M, Quackenboss JJ, Lebowitz MD (1990) Chronic respiratory effects of indoor formaldehyde exposure. *Environ Res* 52: 117–125. https://doi.org/10.1016/S0013-9351(05)80247-6.
- 20. Zomorodian M, Tahsildoost M, Hafezi M (2016) Thermal comfort in educational buildings: A review article. *Renewable and Sustainable Energy Revi* 59: 895–906. https://doi.org/10.1016/j.rser.2016.01.033.
- 21. Gaihre S, Semple S, Miller J, et al. (2014) Classroom carbon dioxide concentration, school attendance, and educational attainment. *J Sch Health* 84: 569–574. https://doi.org/10.1111/josh.12183.
- 22. Aglan HA (2003) Predictive model for CO₂ generation and decay in building envelopes. *J Appl Phys* 93: 1287–1290. https://doi.org/10.1063/1.1529992.
- 23. Caruana-Montaldo B, Gleeson K, Zwillich CW (2000) The control of breathing in clinical practice. *Chest* 117: 205–225. https://doi.org/10.1378/ chest.117.1.205.
- 24. Cheng C, Matsukawa T, Sessler DI, et al. (1995) Increasing mean skin temperature linearly reduces the core-temperature thresholds for vasoconstriction and shivering in humans. *J Am Soc Anesthesiol* 82: 1160–1168, https://doi.org/10.1097/00000542-199505000-00011.
- 25. Evans P, Bristow M, Hucklebridge F, et al. (1994) Stress, arousal, cortisol and secretory immunoglobulin A in students undergoing assessment. *Br J Clin* Psychol 33: 575–576, https://doi.org/10.1111/j.2044-8260.1994.tb01154.x.

- 26. Hamilton M, Rackes A, Gurian PL, et al. (2015) Perceptions in the US building industry of the benefits and costs of improving indoor air quality. *Indoor Air* 26: 318–330. https://doi.org/10.1111/ina.12192.
- 27. H€oppe P, Oohori T, Berglund L, et al. (2000) Vapor resistance of clothing and its effect on human response during and after exercise, in: Proc. CLIMA, 2000, pp. 97–101, 1985.
- 28. Jaggs M, Palmer J (2000) Energy performance indoor environmental quality retrofit—a European diagnosis and decision making method for building refurbishment. *Energy Build* 31: 97–101. https://doi.org/10.1016/S0378-7788(99)00023-7.
- 29. Kim J, Hong T, Jeong J, et al. (2016) An optimization model for selecting the optimal green systems by considering the thermal comfort and energy consumption, *Appl Energy* 169: 682–695. https://doi.org/10.1016/j.apenergy.2016.02.032.
- 30. McCunney RJ (2001), Health and productivity. *J Occup Environ Med* 43: 30–34. https://doi.org/10.1097/00043764-200101000-00007.
- 31. Reynolds SJ, Black DW, Borin SS, et al, (2001) Indoor environmental quality in six commercial office buildings in the Midwest United States. *Appl Occup Environ Hyg* 16: 1065–1077. https://doi.org/10.1080/104732201753214170.
- 32. Varjo J, Hongisto V, Haapakangas A, et al. (2015) Simultaneous effects of irrelevant speech, temperature and ventilation rate on performance and satisfaction in open-plan offices. *J Environ Psychol* 44: 16–33. https://doi.org/10.1016/j.jenvp.2015.08.001.
- 33. Wood TM, Bhat KM (1988) Methods for measuring cellulase activities. *Methods Enzymol* 160: 87–112. https://doi.org/10.1016/0076-6879(88)60109-1.
- 34. ASHRAE, Standard 62.1. 2013. Ventilation for Acceptable Indoor Air Quality, American Society of Heating, Refrigerating and Air-Conditioning Engineers, Inc, Atlanta, GA, 2013.
- 35. Daisey JM, Angell WJ, Apte MG (2003) Indoor air quality, ventilation and health symptoms in schools: an analysis of existing information. *Indoor Air* 13: 53–64. https://doi.org/10.1034/j.1600-0668.2003.00153.x.
- 36. Kajtar L, Herczeg L. (2012) Influence of carbon-dioxide concentration on human well-being and intensity of mental work. *QJ Hung Meteorol Serv* 116: 145–169.
- 37. Kim J, de Dear R (2012) Nonlinear relationships between individual IEQ factors and overall workspace satisfaction. *Build Environ* 49: 33–40. https://doi.org/10.1016/j.buildenv.2011.09.022.
- 38. Muthuraj K, Othmani C, Krause R, et al. (2024) A convolutional neural network to control sound level for air conditioning units in four different classroom conditions (2024). *Energy Build* 324: 114913 https://doi.org/10.1016/j.enbuild.2024.114913
- 39. Othmani C, Sebastian Merchel S, Altinsoy ME, et al. (2023) Acoustic Travel-Time TOMography Technique to Reconstruct the Indoor Temperature: How to Improve the Field Reconstruction Quality? *IEEE T Instrum Meas* 73 https://doi.org/10.1109/TIM.2023.3335531
- 40. Othmani C, Dokhanchi NS, Merchel S, et al. (2023) Acoustic tomographic reconstruction of temperature and flow fields with focus on atmosphere and enclosed spaces: A review. *Appl Therm Eng* 223: 119953. https://doi.org/10.1016/j.applthermaleng.2022.119953
- 41. Gauderman WJ, Vora H, McConnell R, et al. (2007) Effect of exposure to traffic on lung development from 10 to 18 years of age: a cohort study. *Lancet* 369: 571–577. https://doi.org/10.1016/S0140-6736(07)60037-3.

- 42. Fracchia L, Pietronave S, Rinaldi M, et al. (2006). The assessment of airborne bacterial contamination in three composting plants revealed site related biological hazard and seasonal variations. *J Appl Microbiol* 100: 973–984.
- 43. Gorny R L, Reponen T, Willeke k, et al. (2002). Fungal fragments as indoor air biocontaminnts. *Appl Enviro Microbiol* 68: 3522–3531.
- 44. Park DU, Yeom JK, Lee WJ, et al. (2013) Assessment of the levels of airborne bacteria, Gram-negative bacteria, and fungi in hospital lobbies. *Int J Environ Res Public Health* 10: 541–555. https://doi.org/10.3390/ijerph10020541.
- 45. Stryjakowska-Sekulska M, Piotraszewska-Pająk A, Szyszka A, et al. (2007) Microbiological Quality of Indoor Air in University Rooms. *Pol J Environ Stud* 16: 623–632.
- 46. Gołofit-Szymczak M, Górny RL (2010) Bacterial and fungal aerosols in air-conditioned office buildings in Warsaw, Poland--the winter season. *Int J Occup Saf Ergon*. 16: 465–476. https://doi.org/10.1080/10803548.2010.11076861.
- 47. Diriba L, Kassaye A, Yared M (2016) Identification, Characterization and Antibiotic Susceptibility of Indoor Airborne Bacteria in Selected Wards of Hawassa University Teaching and Referral Hospital, South Ethiopia. *OALib* 01: 1–12. http://dx.doi.org/10.4236/oalib.preprints.1200012.
- 48. Fujiyoshi S, Tanaka D, Maruyama F (2017) Transmission of Airborne Bacteria across Built Environments and Its Measurement Standards: A Review. *Front Microbiol* 8: 2336. https://doi.org/10.3389/fmicb.2017.02336.
- 49. Fabian P, Miller S, Reponen T, et al. (2005) Ambient bioaerosol indices for air quality assessments of flood reclamation. *J Aerosol Sci* 36: 763–783. https://doi.org/10.1016/j.jaerosci.2004.11.018.
- 50. Fang X, Zhou Y, Yang Y, et al. (2007) Prevalence and risk factors of trichomoniasis, bacterial vaginosis, and candidiasis for married women of child-bearing age in rural Shandong. *Jpn J Infect Dis* 60: 257–261.
- 51. Frank W Raumklima, thermische Behaglichkeit (1975) Literaturauswertung durchgeführt im auftrage des Bundesministers für Raumordnung, Bauwesen und Städtebau. Berlin: *W Ernst*
- 52. Klindworth A, Pruesse E, Schweer T, et al. (2013) Evaluation of general 16S ribosomal RNA gene PCR primers for classical and next-generation sequencing-based diversity studies. *Nucleic Acids Res* 41: e1. https://doi.org/10.1093/nar/gks808.
- 53. Minot S, Krumm N, Greenfield N (2015) A Sensitive and Accurate Data Platform for Genomic Microbial Identification. *bioRxiv* https://doi.org/10.1101/027607.
- 54. Siegwald L, Touzet H, Lemoine Y, et al. (2017) Assessment of Common and Emerging Bioinformatics Pipelines for Targeted Metagenomics. *PLoS One* 12: e0169563. https://doi.org/10.1371/journal.pone.0169563.
- 55. ANSI/ASHRAE (2001) Ventilation for acceptable indoor air quality: *American Society of Heating, Refrigerating and Air-Conditioning Engineers*.
- 56. Hoffmann B, Boogaard H, de Nazelle A, et al. (2021) WHO Air Quality Guidelines 2021-Aiming for Healthier Air for all: A Joint Statement by Medical, Public Health, Scientific Societies and Patient Representative Organisations. *Int J Public Health* 66: 1604465. https://doi.org/10.3389/ijph.2021.1604465.

- 57. Danger R, Moiteaux Q, Feseha Y, et al. (2021) A web application for regular laboratory data analyses. *PLOS ONE* 16: e0261083. https://doi.org/10.1371/journal.pone.0261083.
- 58. Soares S, Fraga S, Delgado JM, et al. (2015) Influence of Indoor Hygrothermal Conditions on Human Quality of Life in Social Housing. *J Public Health Res* 4: 589. https://doi.org/10.4081/jphr.2015.589.
- 59. He Y, Luo Q, Ge P, et al. (2018) Review on Mould Contamination and Hygrothermal Effect in Indoor Environment. *J Environ Prot* 09: 100–110. https://doi.org/10.4236/jep.2018.92008.
- 60. Wolkoff P, Azuma K, Carrer P (2021) Health, work performance, and risk of infection in office-like environments: The role of indoor temperature, air humidity, and ventilation. *Int J Hyg Environ Health* 233: 113709. https://doi.org/10.1016/j.ijheh.2021.113709.
- 61. Arar M, Jung C (2021) Improving the Indoor Air Quality in Nursery Buildings in United Arab Emirates. *Int J Environ Res Public Health* 18: 12091. https://doi.org/10.3390/ijerph182212091.
- 62. Branco PT, Alvim-Ferraz MC, Martins FG, et al. (2015) Children's exposure to indoor air in urban nurseries-part I: CO₂ and comfort assessment. *Environ Res* 140: 1–9. https://doi.org/10.1016/j.envres.2015.03.007.
- 63. Branco P, Alvim-Ferraz MCM, Martins FG, et al. (2019) Quantifying indoor air quality determinants in urban and rural nursery and primary schools. *Environ Res* 176: 108534. https://doi.org/10.1016/j.envres.2019.108534.
- 64. Mečiarová Ľ, Vilčeková S, Krídlová Burdová E, et al. (2018) The real and subjective indoor environmental quality in schools. *Int J Environ Health Res* 28: 102–123. https://doi.org/10.1080/09603123.2018.1429579.
- 65. Ruggieri S, Longo V, Perrino C, et al. (2019) Indoor air quality in schools of a highly polluted south Mediterranean area. *Indoor Air* 29: 276–290. https://doi.org/10.1111/ina.12529.
- 66. Sá JP, Branco P, Alvim-Ferraz MCM, et al. (2017) Evaluation of Low-Cost Mitigation Measures Implemented to Improve Air Quality in Nursery and Primary Schools. *Int J Environ Res Public Health* 14: 585. https://doi.org/10.3390/ijerph14060585.
- 67. Vassella CC, Koch J, Henzi A, et al. (2021) From spontaneous to strategic natural window ventilation: Improving indoor air quality in Swiss schools. *Int J Hyg Environ Health* 234: 113746. https://doi.org/10.1016/j.ijheh.2021.113746.
- 68. Villanueva F, Notario A, Cabañas B, et al. (2021) Assessment of CO(2) and aerosol (PM(2.5), PM(10), UFP) concentrations during the reopening of schools in the COVID-19 pandemic: The case of a metropolitan area in Central-Southern Spain. *Environ Res* 197: 111092. https://doi.org/10.1016/j.envres.2021.111092.
- 69. Simoni M, Annesi-Maesano I, Sigsgaard T, et al. (2010) School air quality related to dry cough, rhinitis and nasal patency in children. *Eur Respir J* 35: 742–749. https://doi.org/10.1183/09031936.00016309.
- 70. Michelot N, Marchand C, Ramalho O, et al. Monitoring indoor air quality in French schools and day-care centres. Results from the first phase of a pilot survey. Healthy Buildings 2012, 10th International Conference; 2012 2012-07-08; Brisbane, Australia https://hal.archives-ouvertes.fr/hal-00747458/document, https://hal.archives-ouvertes.fr/hal-00747458/file/2012-286_post-print.pdf.

- 71. Clausen G, Toftum J, Andersen B. Indeklima i klasseværelser resultater af Masseeksperiment 2014. (2014) Indoor Environment in Classrooms Results of the Mass Experiment.
- 72. Hanoune B, Carteret M (2015) Impact of kerosene space heaters on indoor air quality. *Chemosphere* 134: 581–587. https://doi.org/10.1016/j.chemosphere.2014.10.083.
- 73. Brdarić D, Capak K, Gvozdić V, et al. (2019) Indoor carbon dioxide concentrations in Croatian elementary school classrooms during the heating season. *Arh Hig Rada Toksikol* 70: 296–302. https://doi.org/10.2478/aiht-2019-70-3343.
- 74. Jovanović M, Vučićević B, Turanjanin V, et al. (2014) Investigation of indoor and outdoor air quality of the classrooms at a school in Serbia. *Energy* 77: 42–48. https://doi.org/10.1016/j.energy.2014.03.080.
- 75. Yuan WM, Lu YQ, Wei Z, et al. (2016) An Epistaxis Emergency Associated with Multiple Pollutants in Elementary Students. Biomedical and environmental sciences: BES. 29: 893–897. http://www.besjournal.com/Articles/Archive/2016/No12/201701/t20170112_137315.html.
- 76. Neamtiu IA, Lin S, Chen M, et al. (2019) Assessment of formaldehyde levels in relation to respiratory and allergic symptoms in children from Alba County schools, Romania. *Environ Monit Assess* 191: 591. https://doi.org/10.1007/s10661-019-7768-6.
- 77. Zhu YD, Li X, Fan L, et al. (2021) Indoor air quality in the primary school of China-results from CIEHS 2018 study. *Environ Pollut* 291: 118094. https://doi.org/10.1016/j.envpol.2021.118094.
- 78. Santos Catai AS, Petruci JFS, Cardoso AA (2024) Compact colorimetric method for determining formaldehyde in indoor air: Applying to an environment contaminated with hand sanitizer vapor. *Build Environ* 257: 111546. https://doi.org/10.1016/j.buildenv.2024.111546.
- 79. Matic B, Rakic U, Jovanovic V, et al. (2017) Key Factors Determining Indoor Air PM(10) Concentrations in Naturally Ventilated Primary Schools in Belgrade, Serbia. *Zdr Varst* 56: 227–235. https://doi.org/10.1515/sjph-2017-0031.
- 80. Branis M, Rezácová P, Domasová M (2005) The effect of outdoor air and indoor human activity on mass concentrations of PM(10), PM(2.5), and PM(1) in a classroom. *Environ Res* 99: 143–149. https://doi.org/10.1016/j.envres.2004.12.001.
- 81. Tran D, Alleman L, Coddeville P, et al. (2012) Elemental characterization and source identification of size resolved atmospheric particles in French classrooms. *Atmos Environ* 54: 250–259. https://doi.org/10.1016/j.atmosenv.2012.02.021.
- 82. Oliveira M, Slezakova K, Delerue-Matos C, et al. (2019) Children environmental exposure to particulate matter and polycyclic aromatic hydrocarbons and biomonitoring in school environments: A review on indoor and outdoor exposure levels, major sources and health impacts. *Environ Int* 124: 180–204. https://doi.org/10.1016/j.envint.2018.12.052.
- 83. Moriske HJ, Drews M, Ebert G, et al. (1996) Indoor air pollution by different heating systems: coal burning, open fireplace and central heating. *Toxicol Lett* 88: 349–354. https://doi.org/10.1016/0378-4274(96)03760-5.
- 84. Ruiz PA, Toro C, Cáceres J, et al. (2010) Effect of gas and kerosene space heaters on indoor air quality: a study in homes of Santiago, Chile. *J Air Waste Manag Assoc* 60: 98–108. https://doi.org/10.3155/1047-3289.60.1.98.
- 85. Hong B, Qin H, Jiang R, et al. (2018) How Outdoor Trees Affect Indoor Particulate Matter Dispersion: CFD Simulations in a Naturally Ventilated Auditorium. *Int J Environ Res Public Health* 15: 2862. https://doi.org/10.3390/ijerph15122862.

- 86. Reimer LC, Sardà Carbasse J, Koblitz J, et al. (2022) the knowledge base for standardized bacterial and archaeal data. *Nucleic Acids Res* 50: D741–D746. https://doi.org/10.1093/nar/gkab961.
- 87. Findley K, Oh J, Yang J, et al. (2013) Topographic diversity of fungal and bacterial communities in human skin. *Nature* 498: 367–370. https://doi.org/10.1038/nature12171.
- 88. Madsen AM, Moslehi-Jenabian S, Islam MZ, et al. (2018) Concentrations of *Staphylococcus* species in indoor air as associated with other bacteria, season, relative humidity, air change rate, and *S. aureus*-positive occupants. *Environ Res* 160: 282–291. https://doi.org/10.1016/j.envres.2017.10.001.



© 2025 the Author(s), licensee AIMS Press. This is an open access article distributed under the terms of the Creative Commons Attribution License (https://creativecommons.org/licenses/by/4.0)