



Research article

Selective adhesion of wastewater bacteria to *Pleurotus ostreatus* mycelium in a trickle-bed bioreactor

Kateřina Svobodová¹, Denisa Petráčková², Hana Szabad¹ and Čeněk Novotný^{1,*}

¹ Laboratory of Environmental Biotechnology, Institute of Microbiology of the CAS, v.v.i., Vídeňská 1083, 14220 Prague, Czech Republic

² Laboratory of Cell Signalization, Institute of Microbiology of the CAS, v.v.i., Vídeňská 1083, 14220 Prague, Czech Republic

* **Correspondence:** Email: novotny@biomed.cas.cz; Tel: +420-29-644-2767.

Abstract: The work is focused on spontaneous colonization of fungal mycelium by invading microorganisms in a trickle-bed fungal bioreactor operating under semi-sterile conditions. *Pleurotus ostreatus* was grown under the flow of synthetic wastewater containing activated sludge bacteria and the microbial consortium developed in the reactor was characterized. Genotype and phenotype profile of the reactor-invading, bacterial consortium was clearly distinctive from that of the original activated sludge. The bacterial consortium from the reactor contained a higher portion of bacteria capable of cellobiose utilization and a small amount of bacteria with the ability to utilize benzoic acids. The invading bacteria had no effect on the dye decolorization performance of the fungal reactor. Five bacterial strains colonizing *P. ostreatus* reactor cultures were isolated and identified as species of the genera *Pseudomonas* and *Bacillus*. Except for *Bacillus cereus* all strains displayed a potential to inhibit fungal growth on solid media (14 to 51 % inhibition) which was comparable or higher than that of the reference bacterial strains. The pH- and media composition-dependence of the growth inhibition was demonstrated.

Keywords: *Pleurotus ostreatus*; activated sludge; fungal/bacterial co-culture; trickle-bed bioreactor; dye decolorization capacity; microbial community structure; fungal growth inhibition

Abbreviations

PAHs: polycyclic aromatic hydrocarbons;

WWTP: wastewater treatment plant;

MEG: malt-extract glucose medium;
PDA: potato dextrose agar;
SWW: synthetic wastewater;
PUF: polyurethane foam;
ABTS: 2,2'-azino-bis(3-ethylbenzothiazoline-6-sulphonic acid);
RBBR: Remazol Brilliant Blue R;
DGGE: denaturing gradient gel electrophoresis;
MALDI-TOF: matrix-assisted laser desorption/ionization-time of flight.

1. Introduction

Biodegradation potential of white rot fungi was well established in previous two decades and various bioreactors based on fungal degradative power have recently been tested for biodegradation of persistent organic pollutants including synthetic dyes, chlorinated compounds, PAHs, polybrominated flame retardants, or various pharmaceuticals [1-5]. Most of those studies were carried out under aseptic conditions in the laboratory and thus the fungal performance under non-sterile conditions is still poorly understood. Better knowledge of fungal behavior in the conditions of bacterial stress is needed to advance the application of fungal reactors in wastewater treatment process.

Several recent studies investigated the effect of exogenous competing microorganisms on biodegradation performance of fungal cultures in bioreactors. *Trametes versicolor* was shown to efficiently degrade pharmaceuticals under non-sterile conditions in an urban wastewater [6]. *Trametes pubescens* MUT 2400 was successfully used in combination with activated sludge for the treatment of textile wastewaters [7]. However, the biodegradation capacity of fungal bioreactor cultures can be negatively affected by contaminating microorganisms originating from wastewater [8,9]. Lu et al. [10] studied the diversity of contaminating fungi in reactors with *Phanerochaete chrysosporium* during the degradation process. Variations of the invading fungal community were correlated with the changes of the dye decolorization capacity observed. Unfortunately, those studies only demonstrated the resulting changes in the biodegradation efficiency without providing information on the mechanisms involved in mutual interactions of the microorganisms.

In the soil environment, it was shown that various bacteria can recognize a number of fungal species in different morphological states, including hyphae, spores and fruiting bodies (reviewed in Hogan et al. [11]). Cho et al. [12] demonstrated that bacterial communities that colonized *Pleurotus ostreatus* mycelium growing on the solid substrate significantly differed in the composition and numbers of the individual bacterial species in dependence on the developmental stage of the fungus, thus indicating that the age of the fungus can play a key role in the interspecific interactions of *P. ostreatus*. This work focused on the effect of bacterial stress on the biodegradation capacity of *P. ostreatus* and on characterization of the bacterial population spontaneously colonizing *P. ostreatus* mycelium in a trickle-bed reactor that was operated under a flow of a suspension of activated sludge bacteria suspended in the simulated wastewater (SWW). The results clearly demonstrated differences between the phenotype and genotyping profiles of the mycelium-colonizing, bacterial populations and of the original activated-sludge populations.

2. Material and methods

2.1. Microorganisms

The fungal strain *P. ostreatus* 3004 CCBAS 692 was obtained from the Culture Collection of Basidiomycetes of the Institute of Microbiology (CAS, Prague, Czech Republic). It was maintained on agar plates with MEG medium (5 g·L⁻¹ malt extract, 10 g·L⁻¹ glucose and 2 g·L⁻¹ agar) and stored at 4 °C. The activated sludge was collected from a municipal wastewater treatment plant (WWTP, Central Bohemia, Czech Republic). It was aerated and stored at 4 °C up to one week until its use. The isolated bacterial strains were stored in 300 g·L⁻¹ glycerol at -80 °C. Bacterial strains *Escherichia coli* CCM3988, *Citrobacter koseri* CCM2535, *Rhodococcus erythropolis* CCM2595, *Serratia marcescens* subsp. *marcescens* CCM303, *Corynebacterium glutamicum* CCM2428 (type strain ATCC13032), *Pseudomonas fluorescens* CCM2115, *Pseudomonas aeruginosa* CCM1960, and *Bacillus subtilis* subsp. *spizizenii* CCM1999 were obtained from the Czech Collection of Microorganisms, Masaryk University (Brno, Czech Republic) and were stored in 300 g·L⁻¹ glycerol at -80 °C.

2.2. Trickle-bed reactors

P. ostreatus was grown in semi-solid-state cultures on 1-cm³ cubes of polyurethane foam (PUF). The cultures (1.5 g of PUF in 50 mL of MEG medium) were inoculated with 10 agar plugs from stock fungal plates and incubated at 30 °C for 12 days. To assemble a reactor, the immobilized fungal mycelium was transferred into a sterilized reactor vessel (glass cylinder, volume 150 mL). Filter-sterilized (0.2 µm) air was delivered to the bottom of the reactor vessel using an aquarium aeration device. The reactor was operated for 24 h at 25 °C using a flow of 3 mL·min⁻¹ of sterile SWW containing activated sludge bacteria (6 × 10⁴ CFU·mL⁻¹). SWW consisted of 2.5 g NaCl, 0.2 g glucose, 1.4 g KH₂PO₄, 0.53 g NH₄Cl in 1 L of distilled water, pH 7.6. The activated sludge bacteria collected in WWTP were centrifuged (5000 g, 20 min, 4 °C), washed with a sterile saline solution and suspended in sterile SWW. Control reactor cultures were grown in SWW without the addition of bacteria.

2.3. Dye decolorization in reactors

Reactors filled with *P. ostreatus* mycelium were first operated for 2 h under a flow (3 mL·min⁻¹) of SWW containing activated sludge bacteria. After the first phase when the fungal reactor was exposed to the activated sludge bacteria, the SWW containing the activated sludge bacteria was replaced with a volume of 300 mL SWW without the bacteria but containing 100 mg·L⁻¹ of Remazol Brilliant Blue R (RBBR, Sigma-Aldrich, Germany) and the reactor was further operated in a circulation mode under a flow of 3 mL·min⁻¹. SWW aliquots (0.5 mL) were collected, centrifuged (1 min, 10,000 g) and the absorbance was measured at 592 nm. All incubations were carried out at 25 °C and in triplicates. In the control reactors no bacteria were added.

2.4. Growth inhibition in agar cultures

Growth inhibition was determined using a dual culture technique by measuring colony growth when the fungus and the bacterium were inoculated opposite to each other on various solid media [13]. Two different modifications of the dual culture technique were used. One technique (cf. Table 2) measured growth inhibition as the width of the inhibition zone between the fungal and bacterial colonies at the time when the growth spontaneously terminated (0, fungal and bacterial colonies overlapped; I, 1–3 mm; II, 4–7 mm; III, ≥ 8 mm), in this case the fungus and the bacterium were inoculated on the agar at the same time. The other technique (cf. Figure 2) determined the growth inhibition by measuring the diameter of the fungal colony on Day 6 after inoculation of the fungus. In this case the bacterium was inoculated on the agar opposite the fungus 48 h after the inoculation of the fungus allowing thus the fungus to resume growth on the agar before the bacterium was inoculated.

Using both techniques the agar plates were inoculated in triplicates with a mycelium-covered disc (diameter 0.9 cm) of a 7-d-old MEG fungal culture and with four 10 μL drops of a bacterial culture (10^{10} CFU $\cdot\text{mL}^{-1}$) placed 1–1.5 cm from the agar plate edge. The bacterial inoculum culture was grown overnight (170 rpm, 30 °C) in Lab Lamco beef extract medium pH 7 or in B1 peptone medium (10 g $\cdot\text{L}^{-1}$ peptic digest of animal tissue, 10 g $\cdot\text{L}^{-1}$ beef extract, and 5 g $\cdot\text{L}^{-1}$ NaCl); the cells were counted by plating. The plates with the fungus and bacteria were incubated at 28 °C and the growth inhibition was measured. Four different agar media were used at pH values of 4.5, 5, 6 and 7 (agar 20 g $\cdot\text{L}^{-1}$): MEG, PDA (39 g $\cdot\text{L}^{-1}$), LB (tryptone 10 g $\cdot\text{L}^{-1}$, yeast extract 5 g $\cdot\text{L}^{-1}$, NaCl 10 g $\cdot\text{L}^{-1}$), 2xTY (tryptone 16 g $\cdot\text{L}^{-1}$, yeast extract 10 g $\cdot\text{L}^{-1}$, NaCl 5 g $\cdot\text{L}^{-1}$).

2.5. Enzyme activity determination

Laccase activity was measured by the oxidation of ABTS [14]. Manganese peroxidase was assayed with 2,6-dimethoxyphenol [15]. The enzyme activities were expressed in units of the enzyme activity (U; 1 unit oxidized 1 μmol of substrate per minute). Laccase associated with fungal mycelium was measured as described by Svobodová et al. [16]. One PUF cube colonized by the fungal mycelium was removed as a sample and the enzyme activity measured was expressed in mU of enzyme activity per one PUF cube.

2.6. CFU determination

At the end of cultivation PUF cubes colonized by fungal mycelium were removed from the reactor vessel and incubated in a volume of 15 mL of SWW on a shaker for 10 min. Then the PUF material was squeezed with sterile tweezers and the SWW containing bacteria was collected. Samples of SWW were serially diluted with sterile saline solution. A volume of 50 μL of the dilutions were plated on B1 peptone containing 1.5% agar and cultivated at 30 °C. After 24 h the bacterial colonies were counted and the CFU concentrations calculated.

2.7. Isolation and identification of bacterial strains

Individual bacterial strains associated with the fungal mycelium were isolated by plating on B1

peptone agar plates and identified by MALDI-TOF mass spectrometry as described by Altun et al. [17]. The species classification was performed by comparing the individual peak lists to the database under control of the MALDI Biotyper program.

2.8. *Biolog EcoPlate assays*

Physiological profiles of the bacterial communities were assayed with Biolog EcoPlates (Biolog, Inc., USA) following the protocols of the manufacturer and the principles given by Garland and Mills [18]. The bacteria from the activated sludge and the bacteria extracted from the reactor cultures were collected by centrifugation (5000 g, 20 min, 4 °C), washed twice with sterile saline solution and suspended in a volume of saline solution to obtain a final bacterial concentration of $OD_{600nm} = 0.3$. A volume of 150 μ L of the bacterial suspension was pipetted to each well in a Biolog EcoPlate. The plates were incubated at 30 °C for 24 h and then the absorbance was recorded at 595 nm. The collected absorbance values were subjected to PCA analysis using BioNumerics v5.00 (Applied Maths).

2.9. *DNA extraction, PCR and DGGE analyses*

DNA was extracted from the pelleted bacterial cells according to Stach et al. [19]. Bacterial rDNA was amplified using 1378R and gc984F primers [20] and 50–100 ng of DNA. PCR cycling conditions were 95 °C for 2 min, followed by six cycles of 95 °C for 45 s, 68–63 °C for 30 s, and 72 °C for 1 min, 30 cycles of 95 °C for 45 s, 55 °C for 30 s, and 72 °C for 1 min, and a final extension step of 72 °C for 20 min.

DGGE analysis of the PCR products was performed with gels containing 8% (w/v) polyacrylamide (37.5:1 acrylamide/bis-acrylamide) with a denaturing gradient of 40–60%. The 100% denaturant corresponded to 7M urea and 40% (v/v) formamide. The electrophoresis was run for 17 h at 85 V and 60 °C in 0.5x TAE buffer. The gels were stained with ethidium bromide and the DGGE band profiles were analyzed by BioNumerics v5.00 (Applied Maths) with the band assignment criteria set to 5% minimum profiling and 2% minimum peak area, relative to the max value of each lane. The resulting band profiles were subjected to PCA analyses performed using the BioNumerics v5.00 software.

3. Results and discussion

3.1. *Characterization of microbial communities in trickle-bed reactors*

P. ostreatus strains are often used in biodegradation studies that explore fungal degradation power in different applications (e.g. [21-23]). However, the advancement of application of fungal cultures in bioremediation of contaminated wastewater requires better understanding of their performance and behavior in non-sterile conditions. In this work, a freshly-grown, *P. ostreatus* mycelium immobilized on PUF in a trickle-bed reactor was exposed to a flow of activated sludge organisms suspended in SWW for 24 h (Table 1, Experiment 1). Before use, the activated sludge removed from WWTP was separated from the activated sludge liquid, washed with saline solution and suspended in sterile SSW to avoid the effects of nutrients and compounds present in the original activated sludge on the activated sludge/fungus interactions. During the period of exposure in the

reactor a 44% drop of the fungal laccase activity was observed (cf. Table 1, Experiment 1). Up to 54.5 ± 37.9 U/L of the laccase activity was eluted from the reactor. No MnP activity was detected in the culture. At the end of the experiment after 24 h, an amount of $2.67E + 08$ of bacteria per mL of the extracting SWW was found to be associated with the reactor PUF carrier (total initial dry mass 1.5 g) after extraction with a total volume of 15 mL SWW. The decrease in laccase activity of the fungus in SWW and the amount of bacteria invading the fungal mycelium were similar to those observed previously in batch cultures of the same fungus [24].

Table 1. Microbial and biochemical characteristics of the reactor under conditions of bacterial stress.

		Experiment 1—bacterial counts and laccase activity after 24h-exposure to activated sludge		Experiment 2—bacterial counts after 24h-exposure to activated sludge and RBBR	
Time (h)	Laccase activity associated with fungal mycelium (mU per PUF cube)	Laccase activity at the reactor outlet ($U \cdot L^{-1}$)	Bacteria detected in PUF carrier ($CFU \cdot mL^{-1}$)	Bacteria detected in PUF carrier ($CFU \cdot mL^{-1}$)	
<i>P. ostreatus</i>					
0	75.2 ± 10.1	13.1 ± 3.7	nd	nd	
24	38.9 ± 3.7	2.89 ± 0.5	nd	nd	
<i>P. ostreatus</i> + activated sludge					
0	75.2 ± 10.1	8.4 ± 0.2	nd	nd	
24	33.4 ± 5.4	54.5 ± 37.9	$2.67E + 08$	$1.44E + 08$	

Not determined: nd.

The bacterial population of activated sludge and the bacteria collected from *P. ostreatus* cultures after a 24h-exposure to activated sludge were subjected to the analysis of the bacterial community structure using DGGE analyses. Amplicons of bacterial 16S rDNA (ca 500bp) were separated using DGGE (Figure 1A) and the obtained DGGE profiles were subjected to PCA analyses (Figure 1B). In the lower part of the profile of the activated sludge bacteria, a diffuse band pattern appeared which may occur when complex microbial communities comprise numerous populations in relatively equivalent proportions [25]. Nevertheless, some bands started to become more prominent in this part of the profile in the case of the bacterial population colonizing the fungal mycelium, thus reflecting the ongoing selection process. Similarly, the PCA analysis (Figure 1B) revealed a substantial alteration of the structure of the bacterial community during the process of colonization with respect to the original bacterial population. The samples of bacteria colonizing *P. ostreatus* mycelium had lower y-coordinates (explaining 21.2% of data variance) and higher x-coordinates (explaining 64.5% of data variance) than the samples of the bacteria of the original activated sludge.

The individual bacterial populations were also compared with respect to their ability to utilize various C-sources using Biolog EcoPlate assays. The bacteria colonizing fungal mycelium showed a lower ability to utilize 2-hydroxy benzoic acid, 4-hydroxy benzoic acid, D-glucosaminic acid, itaconic acid, alpha-D-lactose, and D,L-alpha-glycerol phosphate. On the contrary, they displayed a higher ability to utilize D-cellobiose, glycyl-L-glutamic acid, and L-threonine as C-sources (Figure 1C). Similar to the PCA analyses of the DGGE band patterns (Figure 1A), the PCA analyses

of the bacterial phenotype profiles clearly distinguished the two bacterial populations (Figure 1D). The samples of the bacteria colonizing *P. ostreatus* mycelium had lower y-coordinates (explaining 2.8% of data variance) and higher x-coordinates (explaining 96.5% of data variance).

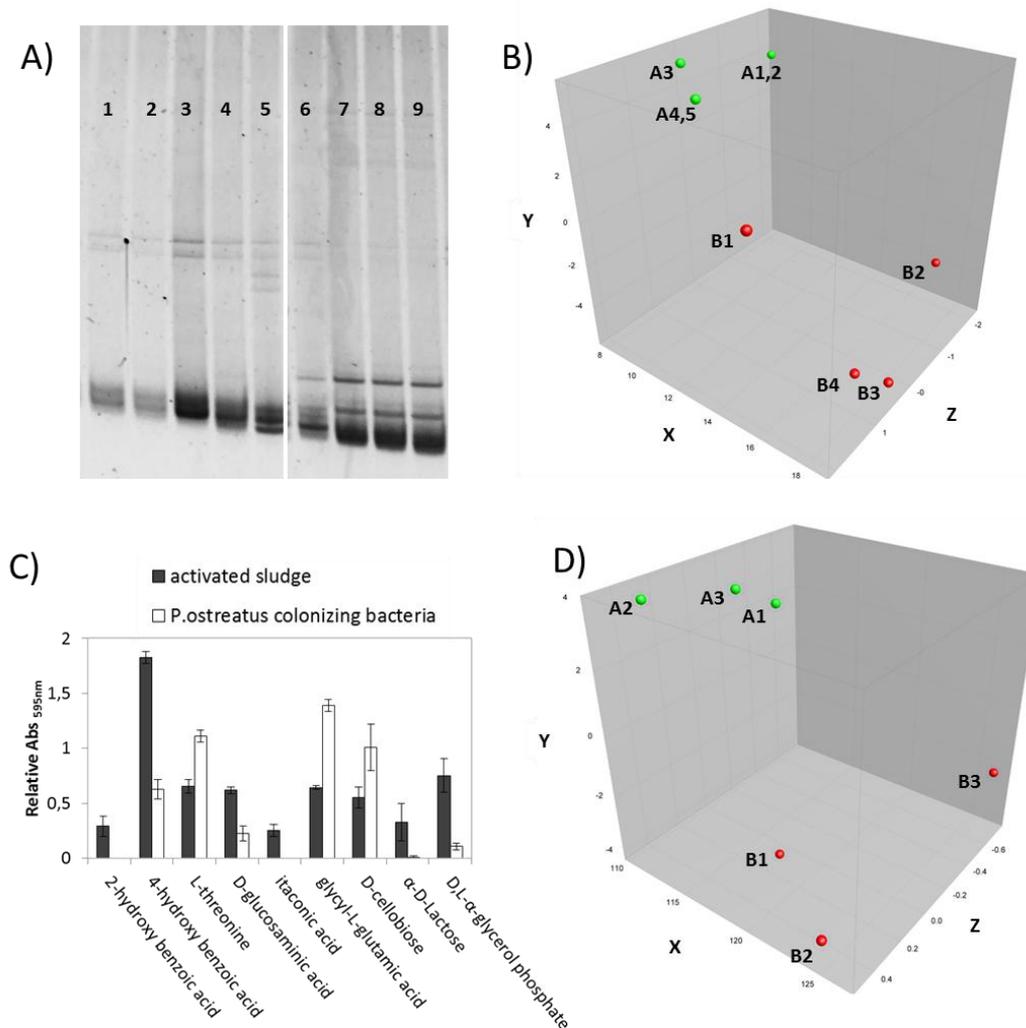


Figure 1. Structures of microbial communities: A) DGGE analyses of bacterial 16S rDNA; lines 1–5: activated sludge bacteria (control); lines 6–9: bacteria colonizing *P. ostreatus* mycelium during 24 h-exposure to activated sludge; B) PCA analyses of DGGE profiles: A-activated sludge bacteria (control); B-bacteria colonizing *P. ostreatus* mycelium during 24h-exposure to activated sludge; C) Utilization of selected C-sources by the bacteria colonizing *P. ostreatus* mycelium during a 24 h-exposure to activated sludge and by those of the original activated sludge (control) as determined by EcoPlate assays; the absorbance detected at 595 nm was related to the average well color development of the assays [18]; D) PCA analyses of phenotype profiles of the bacterial community colonizing *P. ostreatus* mycelium during 24h-exposure to activated sludge; A-activated sludge bacteria (control); B-bacteria colonizing *P. ostreatus* mycelium. PC factors (% of data variance) were: B) x = 64.5%, y = 21.2%, z = 7%; D) x = 96.5%, y = 2.8%, z = 0.3%.

The selection of bacteria during the colonization of *P. ostreatus* mycelium could be driven by the composition of fungal extracellular polysaccharides which show a high variability [26], and could thus reflect the necessity of the presence of corresponding enzyme activities enabling the bacteria to utilize those fungal polysaccharides. The complexity of the substrate was hypothesized to affect the way how bacteria interact by promoting positive interactions, synergistic growth and metabolic activity in bacterial populations in contrast to the situation when glucose was used as the C-source [27]. Our results suggested competition among the bacteria of activated sludge leading to selective colonization of fungal mycelia (Figure 1).

P. ostreatus cultures exposed to the activated sludge bacteria for 2 h were tested for decolorization of RBBR. Under the given conditions the dye decolorization rate reached $40.6 \pm 15.1\%$ after a 24h-incubation, which was comparable to the decolorization capacity of the pure fungal cultures ($49.8 \pm 5.2\%$), showing that no significant change of the decolorization capacity occurred after the exposure to the activated sludge. To summarize, no increase in laccase activity and RBBR decolorization was observed in our mixed cultures, compared to the results of Baldrian [28]. In that work, laccase activity produced by *P. ostreatus* increased significantly after introduction of soil fungi, bacteria and yeast. An increased decolorization of RBBR also occurred in mixed cultures [28].

3.2. Growth inhibition study

General tolerance of *P. ostreatus* 3004 strain to bacteria was tested by dual culture technique with a selection of ubiquitous bacteria by measuring colony growth when the fungus and the bacterium were inoculated opposite to each other on various solid media (pH 4.5–7), covering the optimal pH for both fungal and bacterial growth [13]. The growth inhibition measured by the formation of an inhibition zone between the fungal and bacterial colonies was found with all the bacteria tested on LB and 2xTY media (Table 2). The result probably reflected the metabolic requirements of the partners necessary for growth inhibition that comprised high levels of peptides and amino acids used as nitrogen source. Tryptone-rich media can promote the production of antifungal compounds in bacteria [29]. Nutrient-level effects were reported when growth inhibition was studied with various soil bacteria and *Phanerochaete chrysosporium* [13]. The inhibition with *P. aeruginosa* and *B. subtilis* was also observed on PDA and MEG media, respectively, irrespective of the pH used (Table 2). The behavior of *P. aeruginosa* was similar to that of the agricultural soil pseudomonads towards *P. chrysosporium* described by Radtke et al. [13] where greater antagonism occurred on PDA compared to malt medium. Pseudomonads produce a diverse array of potent antifungal metabolites that include both simple metabolites (phenazin-1-carboxylic acid, pyrrolnitrin, etc.) and complex compounds (e.g. 2,3-de-epoxy-2,3-didehydro-rhizoxin) [30]. The antifungal activity of *B. subtilis* comprises small antibiotic peptides, biosurfactants and fungal wall degrading enzymes. The fact that rich media containing soybean flour used as the nitrogen source support the production of antifungal compounds in *B. subtilis* [29] is in accordance with the inhibition observed on LB, 2xTY and MEG media (Table 2).

There are scarce reports on the antimicrobial properties of *P. ostreatus* [31]. Where no inhibition zones were observed between the partners in our experiments, the bacteria were overgrown with the fungal mycelium after several days. The appearance of the inhibition zones suggested that antifungal compounds were produced by the bacteria that restricted fungal growth. Evidently, *P. ostreatus* was

sensitive to growth inhibition by a number of commonly occurring bacteria, however, this bacterial inhibitory potential could be realized only under rather specific conditions given by nutrient composition of the medium but not the pH value of the medium (cf. Table 2).

MEG agar plates allowing for good growth of both the fungus and the bacteria tested were also used for testing growth inhibition by the strains of the activated sludge bacteria isolated by exposure of the reactor culture of *P. ostreatus* to the activated sludge flow. Five bacterial strains were isolated from those that colonized *P. ostreatus* culture and were designated as AK1, AK2, AK11, AK 38, and AK39. They were identified according to their protein profiles acquired by MALDI-TOF mass spectrometry (data not shown). The strains were classified as *Pseudomonas chlororaphis*, *Pseudomonas veronii*, *Pseudomonas fluorescens*, *Bacillus cereus*, and *B. subtilis*, respectively. Their growth interactions with *P. ostreatus* were compared with five bacteria originating from a culture collection that were also included in the previous testing (cf. Table 2).

Table 2. Growth inhibition of *Pleurotus ostreatus* and various bacteria on different agar media at pH 4.5–7. Growth inhibition was expressed as the width of the inhibition zone between fungal and bacterial colonies (0, fungal and bacterial colonies overlapped; I, 1–3 mm; II, 4–7 mm; III, ≥ 8 mm), the agar media used were malt extract-glucose agar (MEG), potato dextrose agar (PDA), Luria broth (LB), 2xTY medium.

Growth medium pH	Growth inhibition															
	MEG				PDA				LB				2x TY			
	4.5	5	6	7	4.5	5	6	7	4.5	5	6	7	4.5	5	6	7
<i>Escherichia coli</i>	0	0	0	0	-	0	0	0	I	I	II	III	-	III	III	III
<i>Citrobacter koseri</i>	0	0	0	0	-	0	0	0	III	III	III	III	-	III	III	III
<i>Pseudomonas aeruginosa</i>	0	0	0	0	-	I	II	III	I	II	III	III	-	I	III	III
<i>Serratia marcescens</i>	0	0	0	0	-	0	0	0	III	III	0	0	-	III	III	III
<i>Bacillus subtilis</i>	I	II	II	II	-	0	0	0	III	III	III	III	-	II	II	II
<i>Corynebacterium glutamicum</i>	0	0	0	0	-	0	0	0	III	III	III	III	-	II	II	II
<i>Rhodococcus erythropolis</i>	0	0	0	0	-	0	0	0	I	II	III	III	-	III	III	III

In this experiment, the dual culture technique was slightly modified by letting the fungus resume the growth on the agar plate in the course of two days before the bacterium was applied. In this case the inhibition zones were not evaluated but the rate of growth of the fungal colony was determined by measuring its diameter on Day 6 after the inoculation. This approach enabled us to monitor the direct effect on fungal growth by eliminating the effect of a delayed onset of the fungal growth compared to the bacterial one. The fungus overgrew the bacterial colonies in all cases except for *S. marcescens* and *P. chlororaphis* AK1 strains.

With the exception of AK38, all the bacterial strains originating from the activated sludge were able to inhibit the fungal growth to a certain level (Figure 2). The highest inhibition was observed with AK1 and AK39 strains. The growth of *P. ostreatus* in the presence of the collection strains *R. erythropolis*, *C. glutamicum*, *P. fluorescens* and *B. subtilis* was quite comparable to the control (Figure 2). In the case of *S. marcescens* a growth inhibition of ca 40% was observed.

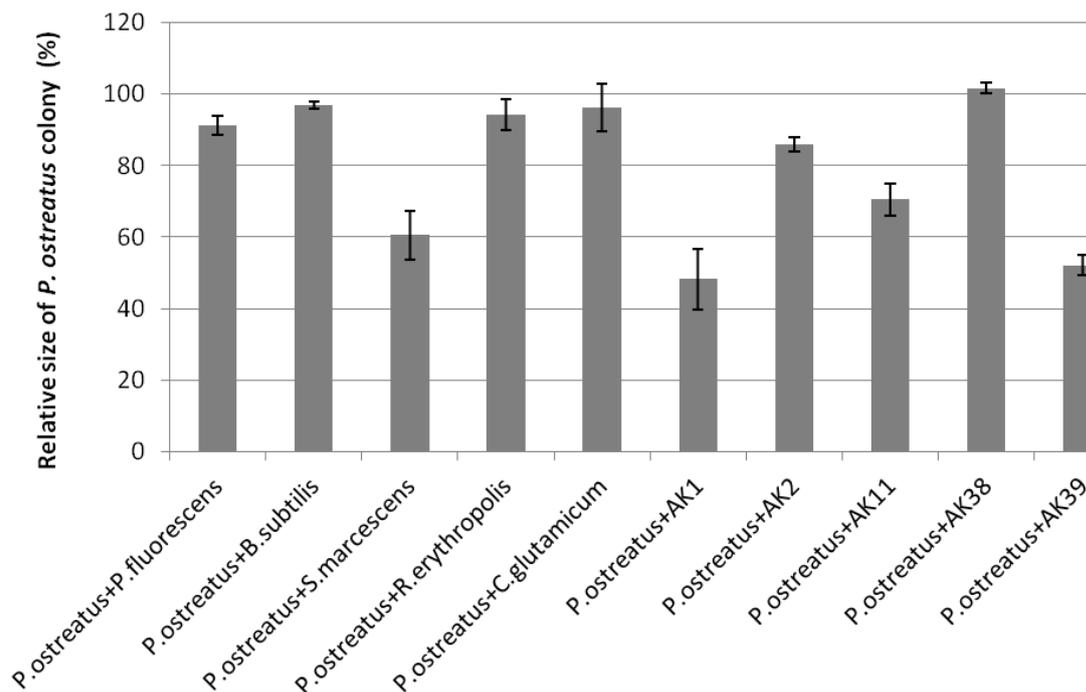


Figure 2. Growth of *P. ostreatus* on MEG agar plates after 6 days of cultivation in the presence of bacteria. Colony diameter of *P. ostreatus* cultures was related to that of *P. ostreatus* monoculture that equalled 100%.

P. chlororaphis UFB2 isolated from a field soil was previously described to possess a broad-spectrum antimicrobial activity connected with multiple gene islands encoding various secondary metabolites [32]. This study shows that its antimicrobial potential also includes *P. ostreatus*. *B. subtilis* and *B. cereus* strains isolated from rhizosphere and roots of oilseed rape exhibited a broad-spectrum antifungal activity towards rice phytopathogenic fungi [33]. *B. subtilis* AK39 isolated from activated sludge inhibited the growth of *P. ostreatus* (Figure 2). The collection strain *B. subtilis* strain CCM1999 did not inhibit fungal growth under those conditions but exhibited an inhibitory effect in the previous test on MEG medium (Table 2). Some strains of fluorescent *Pseudomonas* spp. isolated from commercially-produced *P. ostreatus* were shown to colonize preferentially fungal fruiting bodies and promoted their formation [12]. *P. fluorescens* AK11 was isolated from the *P. ostreatus* mycelium growing in the reactor, however, no primordia/fruiting bodies were observed in the reactor cultures. It had a weak inhibitory effect on the fungal growth but stronger than the collection strain *P. fluorescens* CCM2115 (Figure 2).

4. Conclusions

The analyses of the bacterial community structure showed an ongoing selection process during the colonization of *P. ostreatus* mycelium. The bacteria colonizing fungal mycelium differed from the original population of activated sludge in the community structure and carbon-source utilization potential. The results suggested negative interactions and competition among bacteria of activated sludge leading to selective colonization of fungal mycelia.

Majority of the bacteria originating from both a collection and the activated sludge exhibited an

ability to inhibit fungal growth but this ability strongly depended on the growth conditions and reflected nutrient effects resulting from the composition of the culture medium but not its pH. A high percentage of bacterial strains invading the fungal mycelium were capable to inhibit fungal growth, compared to the collection strains.

The biofilm of *P. ostreatus* mycelium resisted the bacterial invasion and its dye decolorization capacity was not decreased by the colonizing bacteria. It is a precondition for the construction of mixed-culture, trickle-bed reactors and their application to wastewater remediation. The work demonstrated suitability of the *P. ostreatus* trickle-bed reactor for the combined use with activated sludge for wastewater treatment.

Acknowledgement

This work was supported by the Institutional Research Concept RVO 61388971 and was realized in the frame of COST Action ES1403 project. The authors express thanks to E. Sýkorová from the company Hydrotech s.r.o. (Czech Republic) for providing the activated sludge samples.

Conflict of interest

All authors declare no conflicts of interest in this paper.

References

1. Andleeb S, Atiq N, Robson GD, et al. (2012) An investigation of anthraquinone dye biodegradation by immobilized *Aspergillus flavus* in fluidized bed bioreactor. *Env Sci Poll Res* 19: 1728-1737.
2. Rodriguez-Rodriguez CE, Baron E, Gago-Ferrero P, et al. (2012) Removal of pharmaceuticals, polybrominated flame retardants and UV-filters from sludge by the fungus *Trametes versicolor* in bioslurry reactor. *J Haz Mat* 233: 235-243.
3. Rosales E, Perez-Paz A, Vazquez X, et al. (2012) Isolation of novel benzo[a]anthracene-degrading microorganisms and continuous bioremediation in an expanded-bed bioreactor. *Bioprocess Biosys Eng* 35: 851-855.
4. Yadav M, Srivastva N, Shukla AK, et al. (2015) Efficacy of *Aspergillus* sp. for degradation of chlorpyrifos in batch and continuous aerated packed bed bioreactors. *Appl Biochem Biotech* 175: 16-24.
5. Knapp JS, Vantoch-Wood EJ, Zhang F (2008) Use of wood-rotting fungi for the decolorization of dyes and industrial effluents, In: Gadd, G.M. (Ed.) *Fungi in Bioremediation*, Cambridge University Press, 242-304.
6. Cruz-Morato C, Ferrando-Climent L, Rodriguez-Mozaz S, et al. (2013) Degradation of pharmaceuticals in non-sterile urban wastewater by *Trametes versicolor* in a fluidized bed bioreactor. *Water Res* 47: 5200-5210.
7. Spina F, Romagnolo A, Prigione V, et al. (2014) A scaling-up issue: the optimal bioreactor configuration for effective fungal treatment of textile wastewaters. *Chem Eng Trans* 38: 37-42.
8. Gros M, Cruz-Morato C, Marco-Urrea E, et al. (2014) Biodegradation of the X-ray contrast agent iopromide and the fluoroquinolone antibiotic ofloxacin by the white rot fungus *Trametes versicolor*

- in hospital wastewaters and identification of degradation products. *Water Res* 60: 228-241.
9. Li X, de Toledo RA, Wang SP, et al. (2015) Removal of carbamazepine and naproxen by immobilized *Phanerochaete chrysosporium* under non-sterile condition. *New Biotechnol* 32: 282-289.
 10. Lu XM, Ma LH, Wang ZH, et al. (2010) Application of polymerase chain reaction-denaturing gradient gel electrophoresis to resolve taxonomic diversity in white rot fungus reactors. *Env Eng Sci* 27: 493-503.
 11. Hogan DA, Wargo MJ, Beck N, Bacterial biofilm on fungal surfaces, In: Kjelleberg S, Givskov M, eds., *The Biofilm Mode of Life: Mechanisms and Adaptations*, Horizon Scientific Press, 2007, 248.
 12. Cho YS, Kim JS, Crowley DE, et al. (2003) Growth promotion of the edible fungus *Pleurotus ostreatus* by fluorescent pseudomonads. *FEMS Microbiol Lett* 218: 271-276.
 13. Radtke C, Cook WS, Anderson A (1994) Factors affecting antagonism of the growth of *Phanerochaete chrysosporium* by bacteria isolated from soils. *Appl Microbiol Biotechnol* 41: 274-280.
 14. Matsumura E, Yamamoto E, Numata A, et al. (1986) Structure of the laccase-catalyzed oxidation products of hydroxy-benzoic acids in the presence of ABTS (2,2'-azino-di-(3-ethylbenzothiazoline-6-sulfatic acid)). *Agr Biol Chem* 50: 1355-1357.
 15. DeJong E, Cazemier AE, Field JA, et al. (1994) Physiological role of chlorinated aryl alcohols biosynthesized de-novo by the white-rot fungus *Bjerkandera* sp. strain Bos55. *Appl Env Microbiol* 60: 271-277.
 16. Svobodová K, Majcherczyk A, Novotný Č, et al. (2008) Implication of mycelium-associated laccase from *Irpex lacteus* in decolorization of synthetic dyes. *Biores Technol* 99: 463-471.
 17. Altun O, Botero-Kleiven S, Carlsson S, et al. (2015) Rapid identification of bacteria from positive blood culture bottles by MALDI-TOS MS following short-term incubation on solid media. *J Med Microbiol* 64: 1346-1352.
 18. Garland JL, Mills AL (1991) Classification and characterization of heterotrophic microbial communities on the basis of patterns of community-level sole-carbon-source utilization. *Appl Env Microbiol* 57: 2351-2359.
 19. Stach JEM, Bathe S, Clapp JP, et al. (2001) PCR-SSCP comparison of 16S rDNA sequence diversity in soil DNA obtained using different isolation and purification methods. *FEMS Microbiol Ecol* 36: 139-151.
 20. Heuer H, Krsek M, Baker P, et al. (1997) Analysis of actinomycete communities by specific amplification of genes encoding 16S rRNA and gel-electrophoretic separation in denaturing gradients. *Appl Env Microbiol* 63: 3233-3241.
 21. Prabhu MV, Karthikeyan R, Shanmugaprakash M (2016) Modeling and optimization by response surface methodology and neural network-genetic algorithm for decolorization of real textile dye effluent using *Pleurotus ostreatus*: a comparison study. *Desal Wat Treatment* 57: 13005-13019.
 22. Karas PA, Makri S, Papadopoulou ES, et al. (2016) The potential of organic substrates based on mushroom substrate and straw to dissipate fungicides contained in effluents from the fruit-packaging industry—Is there a role for *Pleurotus ostreatus*? *Ecotoxicol Env Safety* 124: 447-454.
 23. Skariyachan S, Prasanna A, Manjunath SP, et al. (2016) Environmental assesment of the

- degradation potential of mushroom fruit bodies of *Pleurotus ostreatus* (Jacq.:Fr.)P.Kumm. towards synthetic azo dyes and contaminating effluents collected from textile industries in Karnataka, India. *Env Monitor Assesment* 188: article nr. 121.
24. Svobodová K, Petráčková D, Kozická B, et al. (2016) Mutual interactions of *Pleurotus ostreatus* with bacteria of activated sludge in solid-bed bioreactors. *World J Microbiol Biotechnol* 32: 94.
 25. Nakatsu CH (2007) Soil microbial community analysis using denaturing gradient gel electrophoresis. *Soil Sci Soc Am J* 71: 562-571.
 26. Ruthes AC, Smiderle FR, Iacomini M (2016) Mushroom heteropolysaccharides: a review on their sources, structure and biological effects. *Carbohydr Polymers* 136: 358-375.
 27. Deng YJ, Wang SY (2016) Synergistic growth in bacteria depends on substrate complexity. *J Microbiol* 54: 23-30.
 28. Baldrian P (2004) Increase of laccase activity during interspecific interactions of white-rot fungi. *FEMS Microbiol Ecol* 50: 245-253.
 29. Kilani-Feki O, Ben Khedher S, Dammak M, et al. (2016) Improvement of antifungal metabolites production by *Bacillus subtilis* V26 for biocontrol of tomato postharvest disease. *Biolog Control* 95: 73-82.
 30. Ligon JM, Hill DS, Hammer PE, et al. (2000) Natural products with antifungal activity from *Pseudomonas* biocontrol bacteria. *Pest Management Sci* 56: 688-695.
 31. Pauliuc I, Botau D (2013) Antibacterial activity of *Pleurotus ostreatus* gemmotherapeutic extract. *J Hort Forest Biotech* 17: 242-245.
 32. Deng P, Wang XQ, Baird SM, et al. (2015) Complete genome of *Pseudomonas chlororaphis* strain UFB2, a soil bacterium with antibacterial activity against bacterial canker pathogen of tomato. *Stand Gen Sci* 10: article No. 117.
 33. Etesami H, Alikhani HA (2016) Rhizosphere and endorhiza of oilseed rape (*Brassica napus* L.) plant harbor bacteria with multifaceted beneficial effects. *Biol Control* 94: 11-24.



AIMS Press

© 2016 Čeněk Novotný et al. licensee AIMS Press. This is an open access article distributed under the terms of the Creative Commons Attribution License (<http://creativecommons.org/licenses/by/4.0>)